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JOSÉ FRANCISCO GONÇALVES JR.\*<sup>, 1</sup>, JULIANA S. FRANÇA<sup>1</sup>, ADRIANA O. MEDEIROS<sup>2</sup>, CARLOS A. ROSA<sup>2</sup> and MARCOS CALLISTO<sup>1</sup>

<sup>1</sup>Universidade Federal de Minas Gerais, Instituto de Ciências Biológicas, Departamento de Biologia Geral, Laboratório de Ecologia de Bentos; e-mail: jfjunior@icb.ufmg.br <sup>2</sup>Departamento de Microbiologia, Laboratório de Ecologia e Biotecnologia de Leveduras, CEP: 31270–901, Belo Horizonte, MG, Brasil

# Leaf Breakdown in a Tropical Stream

key words: invertebrates, bacteria, fungi, yeast, chemical composition.

# Abstract

The objectives of this study were to investigate leaf breakdown in two reaches of different magnitudes, one of a  $3^{rd}$  (closed riparian vegetation) order and the other of a  $4^{th}$  (open riparian vegetation) order, in a tropical stream and to assess the colonization of invertebrates and microorganisms during the processing of detritus. We observed that the detritus in a reach of  $4^{th}$  order decomposed 2.4 times faster than the detritus in a reach of  $3^{rd}$  order, in which, we observed that nitrate concentration and water velocity were greater. This study showed that the chemical composition of detritus does not appear to be important in evaluating leaf breakdown. However, it was shown to be important to biological colonization. The invertebrate community appeared not to have been structured by the decomposition process, but instead by the degradative ecological succession process. With regards to biological colonization, we observed that the density of bacteria in the initial stages was more important while fungi appeared more in the intermediate and final stages.

# 1. Introduction

In streams with well developed riparian vegetation, the study of decomposition of leaf detritus of allochthonous origins is fundamental to understanding energy flow and metabolism maintenance (HENRY *et al.*, 1994; BENFIELD, 1997; WALLACE *et al.*, 1997). Decomposition is characterized by three phases: leaching, conditioning and fragmentation, yet, this sequence should not be considered merely a temporal and successional process, but rather as overlapping and interacting events (GESSNER *et al.*, 1999). The decomposition of leaves is influenced by water characteristics such as the concentration of nutrients (SUBERKROPP and CHAUVET, 1995), temperature (IRONS *et al.*, 1994), water discharge, and pH (WEBSTER and BENFIELD, 1986). But it is also influenced by other factors such as the chemical composition of the detritus (OSTROFSKY, 1997), microorganisms (SUBERKROPP, 1998), and invertebrates (GRACA, 2001).

The microbial community on leaf detritus consists, basically, of fungi and bacteria. In terms of biomass bacteria on leaves are less important than fungi (GRAÇA, 2001). In general, bacteria colonize rapidly during the initial stages of decomposition acting upon easily assimilated molecules (PETERSEN *et al.*, 1989; TANAKA, 1991; JUGNIA *et al.*, 2000). Special attention has been given to fungi, evaluating their role in decomposition and their ability to increase the nutritional value of a substrate (quality of detritus), due to their capacity to

<sup>\*</sup> Corresponding author



metabolize molecules that do not easily decompose such as cellulose and lignin (ABELHO and GRAÇA, 1996; PEREIRA et al., 1998; CANHOTO and GRAÇA, 1999; HAAPALA et al., 2001).

Invertebrates are important in decomposition, mainly in accelerating the process, by fragmenting the detritus into small pieces, thereby increasing the surface area available for microbial action (GRAÇA, 2001). However, the role of these organisms has not yet been determined for tropical streams (MATHURIAU and CHAUVET, 2002; DOBSON *et al.*, 2003).

Headwater streams are strongly influenced by the drainage basins in which they occur. Another important factor is the amount of light that the canopy of the riparian vegetation blocks out. Shading caused by riparian vegetation results in an insufficient amount of light to power autochthonous primary production, causing this type of primary production to be incapable of supporting existing biological communities in these streams. Therefore, the energy to support these communities is often obtained via allochthonous input of organic matter (WEBSTER and MEYER, 1997; JONES, 1997). It is estimated that up to 99% of the energy that circulates in small rivers has its origins from surrounding forests (FISHER and LIKENS, 1973).

The objectives of this study were: (i) to investigate leaf breakdown in two reaches of different magnitudes, one of a 3<sup>rd</sup> (closed riparian vegetation) order and the other of a 4<sup>th</sup> (open riparian vegetation) order in a tropical stream and (ii) to assess the colonization of invertebrates and microorganisms during the processing of detritus.

# 2. Material and Methods

### 2.1. Study Area

The Indaiá stream, part of the Rio Doce watershed, is located in the Serra do Cipó National Park, Minas Gerais, Brazil (PARNA Cipó –  $19^{\circ}20'$  S and  $43^{\circ}44'$  W), in the southern part of the Espinhaço mountain range. The climate in Serra do Cipó is typical for the high altitude tropical with rainy summers and dry winters. Average annual temperatures oscillate between 17.0 and 18.5 °C and the precipitation between 1450 and 1800 mm.

The vegetation in the region is biologically rich with a high rate of endemism (GIULIETTI, 1997). At least three large vegetation formations can be found in the region: tropical savannah (locally called cerrado), rocky fields, and riparian forest.

#### 2.2. Experimental Design

The detritus used in this study, although not identified specifically, was composed of a mixture of recently fallen and commonly encountered senescent leaves from the riparian zone (for an example of vegetation commonly found in this riparian zone see GONÇALVES *et al.*, in press). This study was conducted in two reaches of the Indaiá stream: a  $3^{rd}$  order (closed riparian vegetation,  $19^{\circ}16.4'$  S –  $43^{\circ}31.2'$  W, 1450 m of a.s.l.) and a  $4^{th}$  order (open riparian vegetation,  $19^{\circ}16'$  S –  $43^{\circ}10.9'$  W, 1380 m of a.s.l.).

Leaf detritus was placed in litter bags  $(30 \times 30 \text{ cm})$  with a mesh size of 5 mm. The leaves were weighed (approximately  $6 \pm 0.1$  g of dry weight), After 3, 7, 15, 30, 60, 90, 120, 150, 180, 210, 240 and 270 days of immersion, three litter bags were returned to the laboratory for analysis.

In the field, the following water characteristics were measured with a Horiba model U-10 multiprobe: temperature, pH, conductivity, turbidity and dissolved oxygen. The water velocity was also measured with a flow probe (Global Waters Instrumentation Inc.).

After collection, the content of each bag was washed in running water over a 180  $\mu$ m sieve. The detritus was then dried in a drying oven (60 °C) for 72 h, for later weighing. The breakdown rates were calculated from a negative exponential model (WEBSTER and BENFIELD, 1986). After weighting, the mixed detritus was triturated in order to evaluate the concentrations of Kjeldahl nitrogen (SARRUGE and HAAG, 1974; MALAVOLTA *et al.*, 1989), phosphorous (MIYAZAWA *et al.*, 1992; MALAVOLTA and NETTO 1989), lignin (VAN SOEST, 1963) and carbohydrates (DERIAZ, 1961; CONN and STUMPF, 1975).

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#### 2.3. Microorganisms

For microbiological analysis, one leaf was removed from each sample (litter bag), placed in a sterile bag, and transported to the laboratory on ice to preserve the sample for processing within 8 hours. Each sample was triturated in 2 mL of sterile distilled water using a homogenizer (Polytron PT 2100 Stand Dispersing Equipment) for 30 seconds, previously disinfected with 70% alcohol. To count heterotrophic bacteria 1 mL of the triturated sample was inoculated in triplicate, using the pour plate method, in a standard agar medium – NWRI agar (0.3% soluble peptone; casein; 0.02% K<sub>2</sub>HPO<sub>4</sub>; 0.005% MgSO<sub>4</sub>; 0.0001% FeCl<sub>3</sub>; 1.5% agar). The plates were incubated for 24–48 hours at 25 °C and the results expressed in CFU mL<sup>-1</sup> (CLESCERI *et al.*, 1998). For the isolation of yeasts 0.1 mL of the sample was spread on Ym agar (2% glucose, 1% peptone, 0.3% malt extract, 0.3% yeast extract, 2% agar, 20 mg% chloramphenicol). The plates were incubated for 3-5 days at 25 °C (YARROW, 1998). Colonies were counted and the density of yeasts was estimated by calculating the CFU mL<sup>-1</sup> of the sample. To evaluate the density of the filamentous fungi 1 mL of the sample was inoculated in triplicate, using the pour plate method, in plates containing Martin medium (0.5% peptone, 1% glucose, 0.1% K<sub>2</sub>HPO<sub>4</sub>, 0.05% MgSO<sub>4</sub> · 7 H<sub>2</sub>O, 3.3 mg% Rose bengal, 1.5% agar, pH 5.8) (HORN and DORNER, 1998). The plates were incubated for 3-5 days at 25 °C. Colonies were counted and the results expressed in CFU mL<sup>-1</sup> of the sample.

#### 2.4. Invertebrates

After washing the leaves over a sieve, the remaining material was preserved in 10% formalin for later identification and quantification under a microscope-stereoscope with 30× magnification. The identified invertebrates were deposited into the Benthic Macroinvertebrate Reference Collection of the Instituto de Ciências Biológicas, Universidade Federal de Minas Gerais. The SHANNON-WIENER diversity index was used to assess invertebrate community according to MAGURRAN (1991). Functional feeding groups were classified according to MERRITT and CUMMINS (1996).

#### 2.5. Statistical Analyses

The abundance data matrix, as well as the substrate chemical composition matrix (In transformed), was analyzed by canonical correspondence analysis (CCA) (TER BRAAK, 1986) to determine the relative importance of the environmental variables in the invertebrate community structure.

The analysis of variance (one-way ANOVA) was used to test differences in water parameters within the two reaches. The differences between the decay rates in the two studied reaches were tested with an analysis of covariance analysis (ANCOVA). Multi dependent analysis of variance (MANOVA) was used to test the difference between the chemical composition (nitrogen, phosphorous, lignin and carbohydrate-dependent variables) of the leaf detritus, and the functional feeding groups (abundance of each category-dependent variable) that colonized the detritus. Analysis of variance (two-way ANOVA), was used TO test microorganisms (yeast, bacteria and fungi) and invertebrate (density-dependent variable) that colonized the detritus. Time was an independent variable in all analysis from the two different reaches.

# 3. Results

Results from the water features in the  $3^{rd}$  order (closed riparian vegetation) and  $4^{th}$  order (open riparian vegetation) reach showed that oxygen and depth was significantly greater in the  $4^{th}$  order reach (Table 1). Also of note, the concentrations of PO<sub>4</sub><sup>-</sup>, NH<sub>4</sub>, NO<sub>3</sub>, and velocity values in  $4^{th}$  order reach were higher than in the  $3^{th}$  order.

Leaf breakdown was 2.4 times faster in the 4<sup>th</sup> order reach, when compared to 3<sup>rd</sup> order reach (Fig. 1). In the 4<sup>th</sup> order reach the detritus completely broke down in 240 days (k = 0.0157;  $R^2 = 0.63$ ). In the 3<sup>rd</sup> order reach the detritus breakdown took 270 days after

Parameters	3 <sup>rd</sup> order	4 <sup>th</sup> order	F/p
Depth (cm)	$15.3 \pm 7.8$	$26.5 \pm 7.3$	6.6/0.03
Turbidity (NTU)	13.0	20.0	-
Water Velocity (m s <sup>-1</sup> )	$3.0 \pm 2.3$	$4.9 \pm 3.1$	1.2/0.3
Temperature (°C)	$18.7 \pm 3.1$	$19.4 \pm 4.1$	0.1/0.75
Electrical Conductivity (mS cm <sup>-1</sup> )	$0.012 \pm 0.006$	$0.010 \pm 0.007$	0.47/0.5
$O_2 (mg L^{-1})$	$7 \pm 0.9$	$8.7 \pm 1.4$	7.01/0.02
pH	$5.4 \pm 2.2$	$5.4 \pm 2.1$	0.002/0.9
$NH_4^- (\mu g L^{-1})^*$	48.00	51.32	_
$NO_3 (\mu g L^{-1})^*$	5.193	35.00	_
$PO_4^-$ (µg L <sup>-1</sup> )*	3.057	3.256	_

Table 1.	Water parameters (	mean and	standard	deviation)	in th	he 3 <sup>rd</sup>	and 4th	order	reaches
		of	Indaiá st	ream.					

\* Data for the tropical Cerrado stream were taken from CALLISTO et al. (2004).

incubation, when its remaining weight was 19% of the original dry weight (k = 0.0065;  $R^2 = 0.85$ ). These results were significantly different (ANCOVA, F = 6.17; p = 0.0069).

No variation was observed in the nitrogen concentrations of the detritus of either reach studied (Table 2). The smallest value in the 3<sup>rd</sup> order reach was 1.15% and the largest was 1.36%. In the 4<sup>th</sup> order reach the smallest value obtained was 1.03% and the largest 1.47%. Also no variation was observed in the phosphorous concentrations for the detritus from either the 3<sup>rd</sup> or the 4<sup>th</sup> order reaches throughout breakdown (0.04 to 0.07%). Lignin concentration was less at the start of breakdown (24.3% in the 3<sup>rd</sup> order reach and 20.4% in the 4<sup>th</sup> order reach) and increased, until reaching the maximum of 45.9% in the 3<sup>rd</sup> order reach (on the 120<sup>th</sup> day) and 47.7% (on the 180<sup>th</sup>) day in the 4<sup>th</sup> order reach. Carbohydrate concentrations displayed an opposite pattern. The maximum observed values were obtained at the start of detritus breakdown (20.11% for both reaches). The minimum value of the 3<sup>rd</sup> order reach was 9.18% on the 180<sup>th</sup> and 270<sup>th</sup> days and 9.59% in the 4<sup>th</sup> order on the 180<sup>th</sup> day. Statistical results showed that there was no significant difference in the chemical composition of the detritus between the 3<sup>rd</sup> and 4<sup>th</sup> order reaches for each compound studied (ANOVA, F = 1.94; p > 0.05).



Figure 1. Mass loss of detritus (means of three replicates  $\pm$  standard error) in the 3<sup>rd</sup> (square) and 4<sup>th</sup> (triangle) order reaches of Indaiá stream. ( $\rightarrow$ ) indicate the beginning of the rainy season (September).

Days			С	hemical Co	nposition			
	NITR	OGEN	PHOSE	PHOROUS	LIC	GNIN	CARBOHY	DRATES
	3 <sup>rd</sup>	4 <sup>th</sup>						
1	1.33	1.3	0.07	0.06	24.30	20.40	20.11	20.11
2	1.34	1.25	0.06	0.06	21.20	21.70	13.90	19.48
3	1.30	1.12	0.06	0.05	22.70	24.50	21.29	5.31
10	1.28	1.03	0.06	0.04	25.70	27.30	18.27	17.68
20	1.26	1.26	0.05	0.05	31.30	31.20	15.71	18.09
30	1.25	1.17	0.04	0.05	33.20	31.40	12.33	18.90
60	1.30	1.12	0.05	0.04	38.50	28.70	12.51	11.74
90	1.25	1.20	0.04	0.05	43.20	41.20	11.16	10.57
120	1.36	1.19	0.05	0.05	45.90	44.50	11.74	10.57
150	1.15	1.06	0.04	0.05	45.00	47.70	15.88	10.17
180	1.18	1.25	0.05	0.05	44.70	47.20	9.18	9.59
210	1.19	1.47	0.04	0.07	45.60	43.60	11.34	ND
240	1.16		0.04		45.60		10.35	
270	1.15		0.04		42.31		9.18	

Table 2. Chemical composition of detritus (% DW) during breakdown in the 3<sup>rd</sup> and 4<sup>th</sup> order reaches of Indaiá stream. ND = below limit of detection.

Within the invertebrate groups encountered, Chironomidae (Diptera) and Ephemeroptera were present in all the sampling periods, with Chironomidae occurring in the highest densities (Tables 3 and 4), in the 3<sup>rd</sup> and 4<sup>th</sup> order reaches (5795 and 5549 ind 100 g<sup>-1</sup> DW, respectively), although in the 1<sup>st</sup> and 2<sup>nd</sup> days Ephemeroptera were dominant. The same was observed for Cladocera on the 210<sup>th</sup> day; both occurrences were observed in the 3<sup>rd</sup> order reach. Other groups such as Trichoptera, Oligochaeta, Copepoda, Coleoptera, Odonata, and Hemiptera were also important in both reaches studied. The highest density value for all organisms combined was observed in the 3<sup>rd</sup> order reach, 10229 ind 100 g<sup>-1</sup> DW on the 180<sup>th</sup> day, while the highest density in the 4<sup>th</sup> reach was 9676 ind 100 g<sup>-1</sup> DW on the 210<sup>th</sup> day. The maximum value for the Shannon-Wiener diversity index (H = 0.812), for the 3<sup>rd</sup> order reach, was obtained on the 60<sup>th</sup> day. In the 4<sup>th</sup> order reach the maximum value for this diversity index (H = 0.633).

In the incubated detritus from both reaches we observed that the collector invertebrates were the most abundant, followed by the predators, shredders, and scrappers (Tables 3 and 4). The functional feeding groups had a homogenous distribution and their proportions varied little over time (Fig. 2). A significant difference was found between the reaches of  $3^{rd}$  and  $4^{th}$  orders (ANOVA, F = 18.35, p < 0.001).

The amounts of nitrogen and phosphorous in the detritus were correlated to the invertebrate community at the start of breakdown, while the amount of lignin was correlated to the invertebrates in the more advanced stages of breakdown in both of the reaches studied. The amount of carbohydrate was not correlated to the invertebrates at any point during detritus breakdown (Fig. 3).

The density of heterotrophic bacteria on the detritus, in both reaches, was higher during the following days within each stage: initial stage (1<sup>st</sup>, 2<sup>nd</sup>, and 3<sup>rd</sup> days), intermediate stage (60<sup>th</sup> and 90<sup>th</sup> days), and final stage (150<sup>th</sup>, 180<sup>th</sup>, 210<sup>th</sup>, and 240<sup>th</sup> days) of the decomposition process (Fig. 4A). The highest densities were observed in the initial stages, mainly in the 3<sup>rd</sup> order reach. The yeasts were frequent on the 2<sup>nd</sup>, 3<sup>rd</sup>, 10<sup>th</sup>, and 90<sup>th</sup> days, and the highest density was observed in the 4<sup>th</sup> order reach, on the 90<sup>th</sup> day (Fig. 4B). The density of filamentous fungi increased beginning the 20<sup>th</sup> day, in both reaches studied. Highest densities were

and <b>T</b> stand	lard deviat	ion of	indivi	duals	100 g <sup>-1</sup> D Gathering	w) and y-collect	tunction $S = S$	nal teed nredder, 3	ing grou Sc = Scra	ips (P = per).	predator,	FC = F1	ltering-c	collector	, GC=
Taxa	FFG							days							
		-	2	3	10	20	30	60	90	120	150	180	210	240	270
Annelida															
Hirudinea	Р	$0\pm 0$	$0 \pm 0$	$0\pm 0$	$7 \pm 13$	$7 \pm 12$	$8 \pm 14$	$0\pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$
Oligochaeta Arthropoda	GC	$0\pm 0$	$6 \pm 10$	13 ± 22	46±41	663 ± 347	$1145 \pm 709$	728±412	$26 \pm 25$	$0 \pm 0$	131 ± 114	$261 \pm 193$	$0\pm 0$	$0\pm 0$	23 ± 40
Acerine	D	0+0	0+0	0+0	0+0	0+0	0+0	10 + 17	0+0	cc + vc	0+0	148 + 145	66 + 52	87 + 1/3	10+95
Branchiopoda	4	0-0	0 - 0	0-1-0	0 - 0	0 - 0		11 - 61	0 - 0	77 - +7		140 - 140	or 7 00	C+1 - 70	C0 - 64
Cladocera	FC/P	$6 \pm 11$	$12 \pm 22$	$76 \pm 70$	$208 \pm 322$	$988 \pm 433$	$595 \pm 298$	$451 \pm 345$	$0 \pm 0$	$147 \pm 148$	$94 \pm 47$	$846 \pm 666$	$868\pm855$	$49 \pm 86$	$183 \pm 81$
Ostracoda	FC/P	$0\pm 0$	$0 \pm 0$	$0\pm 0$	$7 \pm 13$	$0\pm 0$	$8 \pm 14$	$19 \pm 17$	19	$0\pm 0$	$0\pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$
Copepoda	CF	$6 \pm 11$	$7 \pm 11$	$19 \pm 18$	$29 \pm 33$	$110 \pm 21$	$128 \pm 86$	$443 \pm 187$	$0\pm 0$	$238 \pm 173$	217 ± 191	$800 \pm 664$	$648 \pm 485$	$52 \pm 50$	$69 \pm 119$
Insecta															
Coleoptera Diptera	Sh/GC/Sc/P	$0\pm 0$	$0 \pm 0$	$6 \pm 11$	22 ± 38	0 ± 0	67 土 49	77 ± 46	$17 \pm 29$	$53 \pm 69$	$54 \pm 63$	532 ± 368	721 ± 451	252 ± 182	179 ± 63
Ceratopogonidae	Р	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$7 \pm 13$	$0 \pm 0$	$52 \pm 65$	$0 \pm 0$	$29 \pm 50$	$0 \pm 0$	$50 \pm 43$	$67 \pm 81$	$0 \pm 0$	$0 \pm 0$
Chaoboridae	Ρ	$0\pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$26 \pm 44$	$0\pm 0$	$0\pm 0$	$0\pm 0$
Chironomidae	P/GC	$38 \pm 32$	$87 \pm 19$	$146 \pm 74$	$1837\pm1317$	$1645 \pm 314$	$2364 \pm 987$	$2009\pm510$	$571 \pm 272$	$1120 \pm 65$	$2070 \pm 1127$	$5795 \pm 4738$	$723 \pm 288$	$253 \pm 84$ 1	$608 \pm 486$
Ephemeroptera	GC	$83 \pm 59$	$151 \pm 56$	$76\pm69$	$384 \pm 330$	$145 \pm 158$	$241 \pm 159$	$461 \pm 249$	$150 \pm 125$	$818 \pm 376$	929 ± 797	$975 \pm 250$	$192 \pm 80$	$110\pm115$	$653 \pm 349$
Heteroptera	Р	$0\pm 0$	$12 \pm 22$	$6 \pm 11$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$10 \pm 18$	$0 \pm 0$	$0 \pm 0$	$13 \pm 23$	$122 \pm 150$	$0\pm 0$	$0\pm 0$	$0\pm 0$
Hemiptera	Р	$0\pm 0$	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$0 \pm 0$	$8 \pm 14$	$18 \pm 16$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$121 \pm 154$
Odonata	Р	$6 \pm 11$	$0 \pm 0$	$0 \pm 0$	$29 \pm 50$	$31 \pm 36$	$8 \pm 14$	$48 \pm 15$	$0 \pm 0$	$0 \pm 0$	$34 \pm 31$	$38 \pm 38$	$0\pm 0$	$0\pm 0$	$0\pm 0$
Trichoptera	Sh/FC/GC/Sc/P	$13 \pm 11$	$0 \pm 0$	$45 \pm 30$	$75 \pm 36$	$98 \pm 99$	$114 \pm 79$	$281 \pm 217$	$201 \pm 127$	$170 \pm 90$	$281 \pm 190$	585 ± 748	$314 \pm 196$	$61 \pm 106$	$44 \pm 38$
Hidrozoa															
Hydra	Р	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0 \pm 0$	$106 \pm 62$	$63 \pm 83$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$
Mollusca															
Gastropoda	Sc	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$36 \pm 63$	$64 \pm 92$	$103 \pm 135$	$33 \pm 58$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0 \pm 0$
Nematoda	Ρ	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$23 \pm 23$	$65 \pm 19$	$34 \pm 13$	$67 \pm 7$	$17 \pm 29$	$0 \pm 0$	$0 \pm 0$	$51 \pm 89$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$
Rotifera	Р	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0 \pm 0$	$0 \pm 0$	$13 \pm 23$	$0\pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$
Total Density		152	276	387	2704	3824	4921	4770	666	2600	3836	10229	3599	859	2929
Diversity H'		0.552	0.496	0.716	0.509	0.667	0.679	0.812	0.535	0.634	0.598	0.680	0.787	0.750	0.607

Table 3. Invertebrates community that colonized detritus during breakdown in the 3<sup>rd</sup> order reach of Indaiá stream (mean of three replicates

Leaf Breakdown

s community that colonized detritus during breakdown in the 4 <sup>th</sup> order reach of Indaiá stream (mean of three rep deviation of individuals 100 g <sup>-1</sup> DW) and functional feeding groups (P = predator, FC = Filtering-collect GC = Gathering-collector, S = Shredder, Sc = Scraper).	dave
Invertebrate id ± standard	FFC
Table 4. cates an	Тауа

					,								
Taxa	FFG						days						
		1	2	3	10	20	30	60	90	120	150	180	210
Annelida													
Oligochaeta	GC	$0\pm 0$	$0\pm 0$	$0\pm 0$	$12 \pm 11$	$41 \pm 70$	$10 \pm 17$	$191 \pm 243$	$0 \pm 0$	$228 \pm 395$	$260 \pm 329$	$0\pm 0$	$92 \pm 159$
Arthropoda Arachnida													
Acarina	Р	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$36 \pm 41$	$123 \pm 141$
Branchiopoda													
Cladocera	FC/P	$6 \pm 11$	$0 \pm 0$	$0\pm 0$	$49 \pm 55$	$256\pm177$	$277\pm137$	$110 \pm 65$	$0 \pm 0$	$23 \pm 40$	$101 \pm 133$	$175 \pm 153$	$92 \pm 159$
Copepoda	FC	$0 \pm 0$	$6 \pm 10$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$0 \pm 0$	$0\pm 0$	$0 \pm 0$	$11 \pm 20$	$0 \pm 0$	$19 \pm 33$	$226 \pm 289$
Insecta													
Coleoptera	Sh/GC/Sc/P	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0 \pm 0$	$7 \pm 12$	$44 \pm 12$	$60 \pm 17$	$0 \pm 0$	$0 \pm 0$	$176 \pm 190$	$540 \pm 316$	$1457 \pm 879$
Diptera													
Ceratopogonidae	Р	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$18 \pm 31$	$0\pm 0$	$0 \pm 0$	$11 \pm 20$	$13 \pm 22$	$0 \pm 0$	$0 \pm 0$
Chironomidae	P/GC	$777 \pm 105$	$497 \pm 287$	$84\pm73$	$1270\pm632$	$503 \pm 41$	$939\pm270$	$1057\pm608$	$278\pm233$	$1026\pm680$	$1976 \pm 1298$	$1467 \pm 1071$	$5549 \pm 2149$
Ephemeroptera	GC	$12 \pm 10$	$83 \pm 98$	$17 \pm 16$	$39 \pm 4$	$110\pm52$	$68 \pm 59$	$266 \pm 231$	$197 \pm 228$	$210\pm115$	$654 \pm 270$	$1155 \pm 944$	$1716 \pm 1127$
Heteroptera	Р	$6 \pm 11$	$0 \pm 0$	$0\pm 0$	$6 \pm 11$	$28 \pm 24$	$20 \pm 35$	$11 \pm 19$	$12 \pm 21$	$0 \pm 0$	$0 \pm 0$	$46 \pm 42$	$0 \pm 0$
Odonata	Р	$19 \pm 33$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	$0\pm 0$	$16 \pm 27$	$0 \pm 0$	$0 \pm 0$	$0 \pm 0$	$14 \pm 24$	$19 \pm 33$	$0 \pm 0$
Trichoptera	Sh/FC/GC/Sc/P	$6 \pm 11$	$0 \pm 0$	$0\pm 0$	$25 \pm 9$	$113\pm143$	$147 \pm 51$	$104 \pm 84$	$63 \pm 52$	$106 \pm 128$	$188 \pm 149$	$80 \pm 33$	$421 \pm 364$
Nematoda	Р	$0 \pm 0$	$0 \pm 0$	$0\pm 0$	$0\pm 0$	7 ± 12	$16 \pm 27$	$0\pm 0$	$0 \pm 0$	$0 \pm 0$	$14 \pm 24$	$19 \pm 33$	$0\pm 0$
			202	101	1011	1001	1 555	0011	022	2121		7220	2020
10tal Density		170	090	101	1401	1004	cccl	1/98	000	1010	1600	0000	0/06
Diversity H'		0.136	0.201	0.197	0.191	0.633	0.568	0.570	0.454	0.494	0.571	0.624	0.555



Figure 2. Abundance of functional feeding groups during detritus breakdown in the 3<sup>rd</sup> (A) and 4<sup>th</sup> (B) order reach. P = Predators, FC = Filtering-collectors, GC = Gathering-collectors, S = Shredders, Sc = Scrapers.

observed in the final stages of breakdown on the  $210^{th}$  day in the  $4^{th}$  order and in the  $270^{th}$  day in the  $3^{rd}$  order reach (Fig. 4C). However, differences found in the densities of bacteria, fungi, and, yeasts proved not to be significant between the  $3^{rd}$  and  $4^{th}$  order reaches (ANOVA, F = 0.92 p > 0.05).

# 4. Discussion

### 4.1. Decomposition Process

Based on PETERSEN and CUMMINS (1974) the decomposition coefficient can be classified (rapid,  $k > 0.01 d^{-1}$ ; intermediate, 0.01  $d^{-1} > k > 0.005 d^{-1}$  and slow  $k < 0.005 d^{-1}$ ) in the 4<sup>th</sup>



Figure 3. Ordination of invertebrates colonization in the  $3^{rd}$  (circle) and  $4^{th}$  (square) order reaches during detritus breakdown (days = numbers) according to the chemical composition variables (vectors) using canonical correspondence analysis (CCA).

order reach as rapid and in the 3<sup>rd</sup> order reach as intermediate. For both reaches of the stream the breakdown rates were very similar to those found for other species in temperate regions (ABELHO, 2001) and tropical regions (MATHURIAU and CHAUVET, 2002; MOULTON and MA-GALHÃES, 2003). However, the decay rates found in this study were less than those found for leaves of native African species (DOBSON *et al.*, 2003).

While the initial chemical composition of the detritus is more important for breakdown (HOORENS *et al.*, 2003), some molecules (e.g., proteins and carbohydrates) can facilitate, while others (e.g., lignin and phenols) can constrain this process (SUBERKROPP *et al.*, 1976; OSTROFSKY, 1997; OSONO and TAKEDA, 1999 and 2001). In our study the chemical composition of detritus was not important in evaluating leaf breakdown, thus we incubated detritus in the two reaches studied with the same chemical characteristics.

Contrary to that which was observed in the chemical composition of the detritus, some abiotic parameters of the waters tested differed in the two reaches of the stream. The stream water velocity, in the 4<sup>th</sup> order reach, was 1.5 times higher than in the 3<sup>rd</sup> order, this difference increased with the onset of rains. Stream water velocity influences decomposition through the physical abrasion and oxygenation of water (we found significant differences) which facilitates breakdown of the detritus (WEBSTER and BENFIELD, 1986). The concentration of nitrate in the water was also higher in the 4<sup>th</sup> order reach than in the 3<sup>rd</sup> order reach. The availability of nutrients in the water has a strong influence on microbial colonization, possibly also affecting decomposition (SUBERKROPP and CHAUVET, 1995; ROSEMOND *et al.*, 2002; CHADWICK and HURYN, 2003; GULIS and SUBERKROPP, 2003). Therefore, the environmental differences (physical and chemical) observed here indicate the strong influences that these factors had over the decay rates found in both reaches, as was observed in CASAS and GESSNER (1999) in the Mediterranean.



Figure 4. Densities (means of three replicates ± standard deviation) of Heterotrophic Bacteria (A), Yeast (B) and Filamentous Fungi (C) during detritus breakdown in the 3<sup>rd</sup> (diamond) and 4<sup>th</sup> (square) order reaches of Indaiá stream.

### 4.2. Biological Colonization

Biological influences are also important factors that may affect the rate of decomposition (GESSNER *et al.*, 1999), through the actions of invertebrates (mainly shredders) and microorganisms (fungi). Chironomidae and Ephemeroptera were the main groups of invertebrates present during the entire decomposition process, confirming recent observations carried out in streams (MATHURIAU and CHAUVET, 2002; MOULTON and MAGALHÃES, 2003) and lakes (NESSIMIAN and DE LIMA, 1997) situated in tropical regions. These two groups could be responsible for structuring the entire invertebrate community due to their ability to colonize, independently of quality and/or time of decomposition, different leaf detritus. These results corroborate those obtained in previous studies which used aquatic macrophyte detritus from a coastal lake (GONÇALVES *et al.*, 2000 and 2003). Furthermore, the invertebrate communi-

ty appears not to have been structured by the decomposition process, which would indicate that this community was organized and structured by the degradative ecological succession process (GONÇALVES *et al.*, 2004).

In our study we observed that the chemical composition of the detritus influenced invertebrate colonization. This may be explained by the time it takes each detritus to become more palatable to invertebrates. Furthermore, it could be an indication of the importance of microorganisms in increasing the palatability of leaf detritus for invertebrates (ROSEMOND *et al.*, 1998; GRAÇA *et al*, 2001). Some evidence in our study indicates this effect: the detritus in the 3<sup>rd</sup> order reach showed a higher diversity and density of invertebrates, where the density of heterotrophic bacteria, yeasts, and filamentous fungi tended to also be higher. Shredders tended to be more abundant in the final stages of decomposition, during which an increase in microorganisms was also observed.

The microbial community is of fundamental importance, throughout the world, in releasing the energy stored in detritus (GULIS and SUBERKROPP, 2003) mainly in tropical regions where shredders are rare (Irons et al., 1994; Dobson et al., 2002). According to SRIDHAR and BÄRLOCHER (2000) and SUBERKROPP (2001) the biomass of fungi associated with detritus was higher than that of bacteria during leaf breakdown, indicating that fungi were the main primary decomposers, accounting for up to 17% of the total biomass of all detritus. In the present study, the highest counts of bacteria in the initial stages of leaf breakdown suggest that bacteria are more important during this period. These results contrast with other results in temperate streams. SAMPAIO et al. (2001) found the highest counts of bacteria in advanced stages of breakdown. Other authors showed that bacteria colonization (measured by biomass) have relatively no importance when compared with fungi in the initial detritus breakdown (WEYERS and SUBERKROPP, 1996; JUGNIA et al., 2000; GRAÇA, 2001). Our results suggest that bacteria metabolize the protein and simple sugars of the detritus during the initial stages and that fungi act on the degradation of complex polysaccharides, mainly those present in the advanced stages. We noted that generally, when counts of bacteria decrease, counts of filamentous fungi increase and this can be explained, as suggested by SUBERKROPP and KLUG (1976) and CHAMIER et al. (1984), by the fact that fungi and bacteria compete with each other for nutrients.

Concerning yeast colonization, it was not possible to identify their effective role in leaf breakdown. As described by SAMPAIO *et al.* (2001), there are several references which show the capacity of yeasts in the degradation of the principal constituents of plants, for example: cellulose degradation and xilanase activities. Our results showed fluctuation in the density of yeasts and according to SAMPAIO *et al.* (2001), who found similar results for yeast counts, this group of fungi presumably are opportunistic organisms assimilating substances released during degradation by other fungi or bacteria.

In summary, we found significant difference in breakdown rates between the reaches studied but not in their biological communities. As colonization of invertebrates and microorganisms seem to be different in tropical and temperate streams, and most of the information available is from the temperate zone, more experiments are needed to elucidate the differences in leaf breakdown in these regions.

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### 6. References

- ABELHO, M. and M. A. S. GRAÇA, 1996: Effects of eucalyptus afforestation on leaf litter dynamics and macroinvertebrate community structure of streams in Central Portugal. Hydrobiologia **324**: 195–204.
- ABELHO, M., 2001: From litterfall to breakdown in stream: A Review. The Scientific World 1: 658–680.
- BENFIELD, E. F., 1997: Comparison of litterfall input streams. Stream Organic Matter Budgets. J.N. Am. Benthol. Soc. 16: 104–108.
- CALLISTO, M., M. GOULART, A. O. MEDEIROS, P. MORENO and C. A. ROSA, 2004: Diversity assessment of benthic macroinvertebrates, yeasts, and microbiological indicators gradient in Serra do Cipó, Brazil. – Braz. J. Biol. 64: 743–755.
- CANHOTO, C. and M. A. S. GRAÇA, 1999: Leaf barriers to fungal colonization and shredders (*Tipula lateralis*) consumpsion of decomposition *Eucalyptus globulus*. Microb. Ecol. **37**: 163–172.
- CASAS, J. J. and M. O. GESSNER, 1999: Leaf litter breakdown in a mediterranean stream characterised by travertine precipitation. Freshw. Biol. **41**: 781–793.
- CHADWICK, M. A. and A. D. HURYN, 2003: Effects of a whole-catchment N addition on stream detritus processing. J.N. Am. Benthol. Soc. 22: 194–206.
- CHAMIER, A. C., P. A. DIXON and S. A. ARCHER, 1984: The spatial distribution of fungi on decomposing alder leaves in a fresh water stream. – Oecologia **64**: 92–103.
- CLESCERI, L. S. E., A. E. GREENBERG and A. D. EATON, 1998: Standard Methods for the Examination of Water and Waste Water. 20<sup>ed</sup>. A.P.H.H, Washington, USA.
- CONN, E. E. and P. K. STUMPF, 1975: Introdução à Bioquímica. São Paulo. Edgard Blucher, Brazil.
- DERIAZ, R. E., 1961: Routine analysis of carbohydrate and lignin in herbage. J. Sci. Food Agric. 12: 150–160.
- DOBSON, M., A. MAGANA, J. M. MATHOOKO and F. K. NDEGWA, 2002: Detritivores in Kenyan highland streams: more evidence for the paucity of shredders in the tropics? Freshw. Biol. 47: 909–919.
- DOBSON, M., J. M. MATHOOKO, F. K. NDEGWA and C. M'ERIMBA, 2003: Leaf litter processing rates in a Kenyan highland stream, the Njoro River. Hydrobiologia **519**: 207–210.
- FISHER, W. G. and G. E. LIKENS, 1973: Energy flow in Bear Brook, New Hampshire: an integrative approach to stream ecosystem metabolism. Ecol. Monogr. 43: 421–439.
- GESSNER, M. O., E. CHAUVET and M. DOBSON, 1999: A perspective on leaf litter breakdown in stream. Oikos 85: 377–384.
- GIULIETTI, A. M., 1997: Flora: diversidade, distribuição geográfica e endemismo. *In*: G. W. FERNAN-DES (ed.), Serra do Cipó: ecologia e evolução. Edição Vozes, Rio de Janeiro, Brazil, 84–100.
- GONÇALVES, J. F. JR., F. A. ESTEVES and M. CALLISTO, 2000: Succession and diversity of Chironomidae in detritus of *Typha domingensis* in a coastal lagoon (Parque Nacional da Restinga de Jurubratiba, State of Rio de Janeiro, Brazil). – Verh. Internat. Verein. Limnol. 27: 2374–2377.
- GONÇALVES, J. F. JR., F. A. ESTEVES and M. CALLISTO, 2003: Chironomids colonization in Nymphaea ampla L. detritus during a degradative ecological successional experiment in a Brazilian coastal lagoon. – Acta Limnol. Brasil. 15: 21–27.
- GONÇALVES, J. F. JR., A. M. SANTOS and F. A. ESTEVES, 2004: The influence of the chemical composition of *Typha domingensis* and *Nymphaea ampla* detritus on invertebrate colonization during decomposition in a Brazilian coastal lagoon. – Hydrobiologia 527: 125–137.
- GONÇALVES, J. F. JR., J. S. FRANÇA and M. CALLISTO, 2006: Dynamics of Allochthonous Organic Matter in a Tropical Brazilian Headstream. – Braz. Arch. Tech. Biol., In press.
- GRAÇA, M. A. S., 2001: The role of invertebrates on leaf litter decomposition in stream-a Review. Internat. Rev. Hydrobiol. 86: 383–393.
- GRAÇA, M. A. S., R. C. F. FERREIRA and C. N. COIMBRA, 2001: Litter processing along a stream gradient: the role of invertebrates and decomposers. – J.N. Am. Benthol. Soc. **20**: 408–420.
- GULIS, V. and K. SUBERKROPP, 2003: Interactions between stream fungi and bacteria associated with decomposing leaf litter and different levels of nutrient availability. Aquat. Microb. Ecol. **30**: 149–157.
- HAAPALA, A., T. MUOTKA and A. MARKKOLA, 2001: Breakdown and macroinvertebrate and fungal colonization of alder, birch, and willow leaves in a boreal forest stream. – J.N. Am. Benthol. Soc. 20: 395–407.

- HENRY, R., V. S. UIEDA, A. A. O. AFONSO and R. M. KIKUCHI, 1994: Imput of allochthonous matter and structure of fauna in a Brazilian headstrem. Verh. Internat. Verein. Limnol. 25: 1866–1870.
- HOORENS, B., R. AERTS and M. STROETENGA, 2003: Does initial litter chemistry explain litter mixture effects on decomposition? Oecologia 142: 578–586.
- HORN, B. W. and J. W. DORNER, 1998: Soil populations of *Aspergillus* species from section flavi along a transect through peanut-growing regions of the United States. Mycologia **90**: 767–776.
- IRONS, J. G., M. W. OSWOOD, R. J. STOUT and C. M. PRINGLE, 1994: Latitudinal patterns in leaf litter breakdown: is temperature really important? – Freshw. Biol. 32: 401–411.
- JONES, J. B. JR., 1997: Benthic organic matter storage in streams: influence of detrital import and export, retention mechanisms, and climate. Stream Organic Matter Budgets. J.N. Am. Benthol. Soc. 16: 109–119.
- JUGNIA, L. B., R. D. TADONLÉKÉ, T. SIME-NGANDO and J. DEVAUX, 2000: The microbial food web in the recently flooded sep reservoir: Diel fluctuation in bacterial biomass and metabolic activity in relation to phytoplankton and flagellate grazers. Microb. Ecol. **40**: 317–329.
- MAGURRAN, A. E., 1991: Ecological diversity and its measurement. London, Chapman and Hall, UK.
- MALAVOLTA, E. and A. V. NETTO, 1989: Nutrição mineral, calagem, cessagem e adubação dos citros. Associação Brasileira para Pesquisa do Potassio e do Fosfato, Piracicaba SP, Brazil.
- MATHURIAU, C. and E. CHAUVET, 2002: Breakdown of leaf litter in a neotropical stream. J.N. Am. Benthol. Soc. 21: 384–396.
- MERRITT, R. W. and K. W. CUMMINS, 1996: An introduction to aquatic insects of North America. 3<sup>rd</sup> Ed. Kendall/Hunt Publishing Company, Iowa, USA.
- MIYAZAWA, M., M. A. PAVAN and M. F. DE BLOCH, 1992: Análise química de tecido vegetal. Londrina: Instituto Agronômico do Paraná, Circular 74, Brazil.
- MOULTON, T. P. and S. A. P. MAGALHÃES, 2003: Responses of leaf processing to impact in streams in atlantic rain forest, Rio de Janeiro, Brazil – A test of the biodiversity ecosystem functioning relationship? – Braz. J. Biol. 63: 87–95.
- NESSIMIAN, J. L. and I. H. A. G. DE LIMA, 1997: Colonização de três espécies de macrófitas por macroinvertebrados aquáticos em um brejo no litoral do Estado do Rio de Janeiro. – Acta Limnol. Brasil. 9: 149–163.
- OSONO, T. and H. TAKEDA, 1999: Decomposition ability of interior and surface fungal colonizers of beech leaves with reference to lignin decomposition. Eur. J. Soil Biol. **35**: 51–56.
- OSONO, T. and H. TAKEDA, 2001: Effects of organic chemical quality and mineral nitrogen addition on lignin and holocellulose decomposition of beech leaf litter by *Xylaria* sp. Eur. J. Soil Biol. **37**: 17–23.
- OSTROFSKY, M. L., 1997: Relationship between chemical characteristics of autumn-shed leaves and aquatic processing rates. J.N. Am. Benthol. Soc. 16: 750–759.
- PEREIRA, A. P., M. A. S. GRAÇA and M. MOLLES, 1998: Leaf litter decomposition in relation to litter physico-chemical properties, fungal biomass, arthropod colonization, and geographical origin of plant species. – Pedobiologia 42: 316–327.
- PETERSEN, R. C. JR. and K. W. CUMMINS, 1974: Leaf processing in a woodland stream. Freshw. Biol. 4: 343–368.
- PETERSEN, R. C. JR., K. W. CUMMINS and G. M. WARD, 1989: Microbial and animal processing of detritus in a woodland stream. – Ecol. Monogr. 59: 21–39.
- ROSEMOND, A. D., C. M. PRINGLE and A. RAMIREZ, 1998: Macroconsumer effects on insect detritivores and detritus processing in tropical stream. – Freshw. Biol. 39: 515–523.
- ROSEMOND, A. D., C. M. PRINGLE, A. RAMIREZ, M. J. PAUL and J. L. MEYER, 2002: Landscape variation in phosphorus concentration and effects on detritus-based tropical streams. – Limnol. Oceanog. 47: 278–289.
- SAMPAIO, A., R. CORTES and C. LEÃO, 2001: Invertebrate and microbial colonisation in native and exotic leaf litter species in a mountain stream. Internat. Rev. Hydrobiol. **4–5**: 527–540.
- SARRUGE, J. R. and H. P. HAAG, 1974: Análises químicas em plantas. Piracicaba: USP/ESALQ, Depto. de Química, Brazil.
- SRIDHAR, K. R. and F. BÄRLOCHER, 2000: Initial colonization, nutrients supply, and fungal activity on leaves decaying in streams. Appl. Environ. Microbiol. **66**: 1114–1119.
- SUBERKROPP, K., G. L. GODSHALK and M. J. KLUG, 1976: Changes in the chemical composition of leaves during processing in a woodland stream. – Ecology 57: 720–727.
- SUBERKROPP, K. and M. J. KLUG, 1976: Fungi and bacteria associated with leaves during processing in

a woodland stream. - Ecology 57: 707-719.

- SUBERKROPP, K. and E. CHAUVET, 1995: Regulation of leaf breakdown by fungi in streams: influences of water chemistry. Ecology **76**: 1433–1445.
- SUBERKROPP, K., 1998: Microorganisms and organic matter decomposition. In: R. J. NAIMAN and R. E. BILBY. River ecology and management: lessons from the Pacific Coastal ecoregion. Springer-Verlag, New York, 120–143
- SUBERKROPP, K., 2001: Fungal growth, production, and sporulation during leaf decomposition in two streams. – Appl. Environ. Microbiol. 67: 5063–5068.
- TANAKA, Y., 1991: Microbial decomposition of reed (*Phragmites communis*) leaves in a saline lake. Hydrobiologia **220**: 119–129.
- TER BRAAK, C. J. F., 1986: Canonical correspondence analysis: a new eigenvector technique for multivariate direct gradient analysis. – Ecology 67: 1167–1179.
- VAN SOEST, P. J., 1963: Use of detergents in the analysis of fibrous feeds. II. A rapid method for the determination of fiber and lignin. – J. Assoc. Off. Agron. Chem. 46: 829–835.
- YARROW, D., 1998: Methods for the isolation and identification of yeasts. *In*: KURTZMAN, C. P. and FELL, J. W. The yeasts, a taxonomic study. Amsterdam: Elsevier, Netherlands, 77–100.
- WALLACE, J. B., S. L. EGGERT, J. L. MEYER and J. R. WEBSTER, 1997: Multiple trophic levels of a forest stream linked to terretrial litter imputs. – Science 277: 102–104.
- WEBSTER, J. R. and E. F. BENFIELD, 1986: Vascular plant breakdown in freshwater ecosystems. An. Rev. Ecol. Syst. 17: 567–594.
- WEBSTER J. R. and J. L. MEYER, 1997: Organic matter budgets for streams: a synthesis. Stream Organic Matter Budgets. – J.N. Am. Benthol. Soc. 16: 141–161.
- WEYERS, H. S. and K. SUBERKROPP, 1996: Fungal and bacteria production during the breakdown of yellow poplar leaves in 2 streams. – J.N. Am. Benthol. Soc. 15: 408–420.

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